## RESEARCH



# Isolation of a novel multiple-heavy metal resistant *Lampropedia aestuarii* GYF-1 and investigation of its bioremediation potential

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### Abstract

**Background** Heavy metal contamination has been a severe worldwide environmental issue. For industrial pollutions, heavy metals rarely exist as singular entities. Hence, researches have increasingly focused on the detrimental effect of mixed heavy metal pollution. Genome analysis of *Lampropedia* strains predicted a repertoire of heavy metal resistance genes. However, we are still lack of experimental evidence regarding to heavy metal resistance of *Lampropedia*, and their potential in mixed heavy metal removal remain elusive.

**Results** In this study, a *Lampropedia aestuarii* strain GYF-1 was isolated from soil samples near steel factory. Heavy metal tolerance assay indicated *L. aestuarii* GYF-1 possessed minimal inhibition values of 2 mM, 10 mM, 6 mM, 4 mM, 6 mM, 0.8 mM, and 4 mM for CdCl<sub>2</sub>, K<sub>2</sub>CrO<sub>4</sub>, CuCl<sub>2</sub>, NiCl<sub>2</sub>, Pb(CH<sub>3</sub>COO)<sub>2</sub>, ZnSO<sub>4</sub>, and FeCl<sub>2</sub>, respectively. The biosorption assay demonstrated its potential in soil remediation from mixed heavy metal pollution. Next the draft genome of *L. aestuarii* GYF-1 was obtained and annotated, which revealed strain GYF-1 are abundant in heavy metal resistance genes. Further evaluations on differential gene expressions suggested adaptive mechanisms including increased lipopolysaccharides level and enhanced biofilm formation.

**Conclusion** In this study, we demonstrated a newly isolated *L. aestuarii* GYF-1 exhibited mixed heavy metal resistance, which proven its capability of being a potential candidate strain for industrial biosorption application. Further genome analysis and differential gene expression assay suggest enhanced LPS and biofilm formation contributed to the adaptation of mixed heavy metals.

**Keywords** Lampropedia aesturaii, Mixed heavy metal stress adaptation, Genome analysis, Relative gene expression, Bioremediation

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#### Introduction

Heavy metal contamination has been a severe worldwide environmental issue due to their toxicity, accumulative, and nonbiodegradable properties. The primary cause of heavy metal pollution has emerged from industrial activities, such as mining, electroplating, paints and pigments, batteries, tanning and textile, steel industries. Use of pesticides, insecticides, fertilizers in agricultural fields have been secondary source of heavy metal contamination [1]. Recently, research focuses on metal pollution have trends to shift from single metals to mixed metals. For industrial pollutions, heavy metals rarely exist as singular entities. Hence, researches have increasingly focused on the detrimental effect of mixed heavy metal pollution to ecosystem and health of living system. For example, heavy metal mixture induces global iron starvation resulted in decreased activity of biological nitrate removal [2]. Exposure to mixed heavy metals is negatively associated with renal function via oxidative stress disorder [3]. Accumulation of mixed heavy metals is highly related to the occurrence of cancer [4]. Thus, effective approaches must be taken for remediation of mixed heavy metals contamination.

Microorganisms have developed various mechanisms for adaptation of mixed heavy metal stress, resulting in an eco-friendly and cost-effective strategy called biosorption. The detoxification of mixed heavy metals includes biosorption via production of extracellular polymeric substances (EPS), efflux of toxic metals by active transporters, intracellular sequestration, surface precipitation, metal reduction. To date, many bacterial groups are considered as potential bioagents for mixed heavy metal removal, such as Bacillus sp., Pseudomonas sp., Alcaligenes sp., Rhizopus sp., Sphingomonas sp., Azospira sp. and Cupriavidus sp., etc [5, 6]. Therefore, discovery of novel bacteria contributes to mixed heavy metal removal is of great importance, which would benefit not only to better understanding of the bacterial adaptation mechanisms, but also to development of new biosorption strategies.

Lampropedia spp. is a Gram-negative, Neisser-positive, non-spore forming coccus. L. hyalina was firstly

**Table 1** Measurement of heavy metals in soil sample (n = 10)

Heavy metal	Origin HM conc. (mg/kg soil) n = 10	Stimu- lated heavy metal me- dium (μM)
Cd <sup>2+</sup>	1.12	0.2
Cr <sup>6+</sup>	65.00	25
Cu <sup>2+</sup>	64.00	20
Ni <sup>2+</sup>	44.0175	15
Pb <sup>2+</sup>	167.4	15
Zn <sup>2+</sup>	261.52	80
Fe <sup>2+</sup>	280	100

isolated from polluted water sample by Schroeter in 1886 [7]. Since then, the other three *L. hyalina* strains, ATCC 11,041, ATCC 43,383, DSM 15,336 were identified from dairy farm yard [8], rumen [9], and activated sludge [10], respectively. L. hyalina DSM 15,536 was identified as phosphorus removal bacteria because its capability of synthesizing polyphosphate and polyhydroxyalkanoates accumulating bacteria [10]. This suggest Lampropedia hyalina might also be resistant to heavy metal stress via polyphosphates-mediated detoxification. Later, researches isolated three other species of Lampropedia, including Lampropedia aestuarii YIM MLB12 from a sediment sample of the Maliao River estuary [11], Lampropedia puvangensis 2-bin from cankered bark tissue of *Populus*  $\times$  *euramericana* [12], and *Lampropedia cohae*rens CT6 from arsenic rich microbial mats of a Himalavan hot water spring [13]. To note, the draft genome of *L*. cohaerens CT6 also indicated a repertoire of heavy metal resistance genes against arsenic, copper, cobalt, zinc, magnesium, and cadmium [13]. However, to our knowledge, we are currently lack of experimental evidence regarding to heavy metal resistance of Lampropedia, and their potential in mixed heavy metal removal remain elusive.

The present study aimed to evaluate the heavy metal resistance and application value of *Lampropedia aestuarii* GYF-1 isolated from soil samples near steel industry. Results demonstrated resistances of *L. aestuarii* GYF-1 to  $Cd^{2+}$ ,  $Cr^{6+}$ ,  $Cu^{2+}$ ,  $Ni^{2+}$ ,  $Pb^{2+}$ ,  $Zn^{2+}$  and  $Fe^{2+}$  as well as its capability of removing mixed heavy metal contamination. Further integrated genomic analysis and relative gene expression results elucidate adaptive mechanisms for reducing mixed heavy metal stress.

#### Results

#### Determination of heavy metal in soil samples

We collected total 10 soil samples (0-10 cm) near steel industry in Liuzhou Guangxi Province, where is biggest steel industrial district of south China. The concentrations of heavy metal were evaluated by AAS (Table 1). The results indicated seven heavy metals Cd, Cr, Cu, Ni, Pb, Zn, Fe with average concentration of 1.12, 65.0, 64.0, 44.02, 167.4, 261.52, 280 mg/kg soil, respectively, existed in the samples. We assumed all chromium ions and ferrous ions present in the sample are hexavalent and divalent, respectively, because (1) AAS did not give the valance of information of heavy metals; (2)  $Cr^{6+}$  and  $Fe^{2+}$  are usually more toxic than  $Cr^{3+}$  and  $Fe^{3+}$ . Thus, a stimulated heavy metal medium (SHMM) containing seven heavy metals (0.2 µM CdCl<sub>2</sub>, 25 µM K<sub>2</sub>CrO<sub>4</sub>, 20 μM CuSO<sub>4</sub>, 15 μM NiCl<sub>2</sub>, 15 μM Pb(CH<sub>3</sub>COO)<sub>2</sub>, 80 μM  $ZnSO_4$ , 100  $\mu$ M FeSO<sub>4</sub>) was formulated that mimics the heavy metals that present in the soil digested solution (Table 1).

#### Isolation of strain GYF-1 from soil sample

Screening of heavy metal resistant bacteria was performed on  $0.1 \times$  SHMM agar plate with serially diluted soil suspensions. One of the isolates was identified as Lampropedia spp. GYF-1 by 16 S rDNA with identity of 99.92% and 99.84% to L. hyalina X32 and L. aestuarii YIM MLB12, respectively. GYF-1 colonies were white, circular, opaque, slimy, with ambiguous margin colonies on LB agar (Fig. 1a), and yellow, coarse, wrinkled, with irregular margin on TSA (Fig. 1b), which were different from currently published Lampropedia spp. strains. Microscopic check revealed coccoid, unchained cells (Fig. 1c). Strain GYF-1 can grow at a range of temperature from 25°C to 37°C with optimal of 30°C. It could grow in a pH 6.5-8.0, and optimal at pH 7.0. We next performed phylogenetic analysis to gather information on biological diversity and genetic classifications of Lampropedia spp. GYF-1. The evolutionary relationship of strain GYF-1 was inferred using Neighbor-Joining based on Kimura 2-parameter method [14]. Strain GYF-1 is phylogenetically close to L. hyalina X32, L. aestuarii YIM MLB12, and Lampropedia sp LJH44 (Fig. 2).

Next, we performed genome analysis (Fig. S1-S5 and Table S1a) and followed by stimulated DNA-DNA hybridization (DDH) by using Genome-to-genome Distance Calculator (GGDC) [15] to identify taxonomy of strain GYF-1 (Table S1b). The draft genome of strain GYF-1 was compared to the genomes of four published *Lampropedia* strains. The results showed strain GYF-1 has close DDH similarity of 87.1% to *L. aestuarii* YIM MLB12 and low similarities to *L. cohaerens* CT6, *L. puy-angensis* 2-bin, and *L. hyalina* DSM 16,112, which indicated GFY-1 is a newly discovered *L. aestuarii* strain.

#### Heavy metal resistance of strain GYF-1

The resistance of strain GYF-1 to  $Cd^{2+}$ ,  $Cr^{6+}$ ,  $Cu^{2+}$ ,  $Ni^{2+}$ ,  $Pb^{2+}$ ,  $Zn^{2+}$ , or  $Fe^{2+}$ , respectively, was determined by

minimum inhibition concentration (MIC). The overnight cultured seeds were inoculated to LB medium supplemented with individual heavy metal at  $OD_{600}$  of 0.05, and incubated at 30°C for 48 h. The MIC values suggested that 2 mM of CdCl<sub>2</sub>, 8 mM of K<sub>2</sub>CrO<sub>4</sub>, 6 mM of CuCl<sub>2</sub>, 4 mM of NiCl<sub>2</sub>, 6 mM of Pb(CH<sub>3</sub>COO)<sub>2</sub>, 0.8 mM of ZnSO<sub>4</sub>, or 4 mM FeCl<sub>2</sub>, respectively, is able to inhibit the growth of strain GYF-1 (Table 2).

#### Removal of individual heavy metal

Next, the metal removal capacity of strain GYF-1 was evaluated. The concentration of individual heavy metal was chosen to be lower than MIC value so that ensure the growth of the strain. Therefore, LB medium supplemented with 0.8 mM of CdCl<sub>2</sub>, 2 mM of K<sub>2</sub>CrO<sub>4</sub>, 2 mM of CuCl<sub>2</sub>, 2 mM of NiCl<sub>2</sub>, 2 mM of Pb(CH<sub>3</sub>COO)<sub>2</sub>, 0.6 mM of ZnSO<sub>4</sub>, or 2 mM FeCl<sub>2</sub>, respectively, was selected (Fig. 3). Strain GYF-1 showed fast removal of  $Cd^{2+}$  and  $Cu^{2+}$  with removal rate of 90% and 73%, respectively, in 12 h at a relatively low cell density. However, the residual Cd<sup>2+</sup> and Cu<sup>2+</sup> maintain at 10% and 25%, respectively, in the medium even if the cell growth continued (Fig, 3a and 3c). Strain GYF-1 reached maximum removal rate of 90% at 24 h, but gradually released 10% of Cr<sup>6+</sup> into the medium from 24 to 72 h (Fig. 3b). This might be due to the cell death under chromium stress although cell density remain increase. The residual Pb<sup>2+</sup> and Zn<sup>2+</sup> decreased very fast with 62% and 54% removal, respectively, at first 12 h, and the removal continued until 72 h at a relatively low adsorption rate (Fig. 3e and f). The rapid removal of Fe<sup>2+</sup> ended at 48 h with Fe<sup>2+</sup> concentration of 21% left in the medium, and strain performed a low removal rate from 48 to 72 h even if the cell density reached  $OD_{600}$  of 1.4 at final (Fig. 3g). Surprisingly, strain GYF-1 could not remove Ni<sup>2+</sup> although it grew well as the same in other heavy metals (Fig. 3d), which suggest strain GYF-1 might use efflux system instead of cell



Fig. 1 Cell morphology of *L. aestuarii* GYF-1. Colony morphology of *L. aestuarii* GYF-1 on (**A**) LB agar plate and (**B**) TSA plate; (**C**) microphotograph of *L. aestuarii* GYF-1, bar, 10 μm



Fig. 2 Phylogenetic analysis of *L. aestuarii* GYF-1 based on 16 S rRNA sequences. The 16 S ribosomal RNA sequences were aligned with BioEdit and phylogenetic tree was calculated in MEGA 7.0 by Neighbor-Joining method with bootstrap value of 1000 replicates

surface adsorption or accumulation when nickel ions are present as sole heavy metal stress.

#### Removal of mixed heavy metals

We first evaluated growth of strain GYF-1 in LB medium supplemented with  $1 \times$  SHMM. Both specific growth rate and max OD<sub>600</sub> are impaired in SHMM medium, suggest multiple heavy metal stress (Fig. S6). Considering (1) relatively low cell density of GYF-1 in SHMM medium; and (2) affinity to mixed heavy metals are unknown, we performed two-round sequential removal for soil extract. The first-round removal was able to reduce 98.5%, 98%, 98%, 84.7%, 92%, 99%, and 98% of Cd<sup>2+</sup>, Cr<sup>6+</sup>, Cu<sup>2+</sup>, Ni<sup>2+</sup>, Pb<sup>2+</sup>, Zn<sup>2+</sup> and Fe<sup>2+</sup>, respectively, suggest still high amount of Ni<sup>2+</sup> and Pb<sup>2+</sup> ions left (Fig. 4). After second-round removal, the levels of Cd<sup>2+</sup>, Cr<sup>6+</sup>, Cu<sup>2+</sup>, and Zn<sup>2+</sup> were under detection, and residual concentrations of

Ni<sup>2+</sup>, Pb<sup>2+</sup> and Fe<sup>2+</sup> were in accepted levels, respectively. Interestingly, strain GYF-1 did not show nickel removal when Ni<sup>2+</sup> solely presented in the medium (Fig. 3d); however, GYF-1 was able to remove Ni<sup>2+</sup> ions in mixed heavy metal stress (Fig. 4d). One explanation is that nickel resistance of GYF-1 might efflux mediated but with no adsorption of the nickel ions, which agreed with that many nickel transporters are annotated on the genome (Table S4). On contrary, the EPS triggered by other heavy metals could also interact with Ni<sup>2+</sup> ions in mixed heavy metal condition.

# Lipopolysaccharides contributes to mixed heavy metal stress adaptation

Next, the adaptive mechanisms used by *L. aestuarii* GYF-1 to mixed heavy metal stress were evaluated. Lipopolysaccharides is the major outer surface

	W		oltion concer	itration of st	rain GYF-T II	i Lb meaiur.	ı suppreme	ntea with	indicated I	neavy met.	al (n=5)			
eavy metal	0.01 mM	0.1mM	0.2mM	0.4mM	0.6mM	0.8mM	1mM	2mM	4mM	6mM	8mM	10mM	0.1 M	MIC (mM)
d <sup>2+</sup>	+++++	+++++	++++++	++++	+++++++++++++++++++++++++++++++++++++++	+	+	+	\ \	\ \	\ \	/	/	2
1-0+	++++	++++	+++	++++	++++	++++	+++++	+++++	++++	++	+	/	/	8
:u <sup>2+</sup>	++++	++++	+++	++++	++++	++++	++	+++++	+	+	/	/	/	9
<sup>2+</sup>	++++	++++	+++	++++	++++	++++	+++++	+++++	+	/	/	/	/	4
b <sup>2+</sup>	+++++	++++	+++	++++	++++	++++	+++++	+++++	++	++	/	/	/	9
2 <sup>+</sup>	+++++	++++	+++++++++++++++++++++++++++++++++++++++	++	+	+	/	/	/	/	/	/	/	0.8
e <sup>2+</sup>	++++	++++++	+++++	+++++	+++++	++++++	++++	++++	+	/	/	/	/	4

sequence gaps are missing in the L. aestuarii GYF-1 genome. Gram-negative bacteria usually express LPS consists of O-antigen, complete core oligosaccharides, and the lipid A. Previous research indicated some bacteria modify LPS amount or structure to adaptively respond to metal stress [17]. To examine if L. aestuarii GYF-1 also use this strategy to possess mixed heavy metal adaptation, we first compared LPS level in between control and mixed heavy metal condition (Fig. 5). To examine if L. aestuarii GYF-1 also use this strategy to possess mixed heavy metal adaptation, we first compared LPS level in between control and mixed heavy metal condition (Fig. 5). The silver staining result indicated significantly increased level of LPS in mixed heavy metals, but the band patterns of the LPS were similar in both conditions, suggest mixed heavy metal adaptation was mainly through adjustment of LPS level. Next, we compared relative expression level of genes involve in LPS biosynthesis. As expected, expressions of four genes kdsA (GYF\_00028), kdsB (GYF\_02650), kdsC (GYF\_02059) kdsD (GYF\_02058) for 3-deoxy-D-manno-octulosonic acid (KDO) biosynthesis are upregulated under mixed heavy metal stress (Fig. 6a). To note, GYF\_03746 and GYF\_03747 were predicted as arabinose-5-phosphate isomerase and 3-deoxy-Dmanno-octulosonate 8-phosphate phosphatase, respectively; however, they were not up-regulated under stress condition. For Lipid A biosynthesis, the relative expressions of all detected genes were significantly increased in mixed heavy metal stress (Fig. 6b). The rgaG, rfaS, and kdtA involved in core oligosaccharide biosynthesis were at least 2-fold increased (Fig. 6c). Taken together, these up-regulated gene expressions indicated the requirement of core oligosaccharide and lipid A in mixed heavy metal adaptation.

in most Gram-negative bacteria [16]. Annotations of L. aestuarii GYF-1 genome indicated genes involve in lipopolysaccharide biosynthesis (Table 3). However, genes encoding biosynthesis of an inner core with three L-glycero-D-manno-heptose sugars (HepI, HepII, and HepIII) are missing, which probably because a few

O-antigen is a polymer with highly variable oligosaccharide repeating subunits at the most external portion of the LPS [16]. Annotations of L. aestuarii GYF-1 genome reveals the presence of four genes involve in L-rhamnose biosynthesis which is an important component of O- polysaccharide [18]. We noticed most of genes involved in O-polysaccharide receptor and polysaccharide biosynthesis were not up-regulated under mixed heavy metal stress, except for dTDP-glucose 4,6-dehydratase rfbB (GYF\_02289) (Fig. 6d), suggest the regulation of O-polysaccharide biosynthesis was not mainly at transcriptional level in mixed heavy metal stress.

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Fig. 3 Single heavy metal biosorption of strain GYF-1. The LB medium supplemented with (**A**) 0.8 mM of CdCl<sub>2</sub>, (**B**) 2 mM of K<sub>2</sub>CrO<sub>4</sub>, (**C**) 2 mM of CuCl<sub>2</sub>, (**D**) 2 mM of NiCl<sub>2</sub>, (**E**) 2 mM of Pb(CH<sub>3</sub>COO)<sub>2</sub>, (**F**) 0.6 mM of ZnSO<sub>4</sub>, or (**G**) 2 mM FeCl<sub>2</sub>, respectively, was inoculated with overnight GYF-1 culture broths at OD<sub>600</sub> of 0.05. The cultures were incubated for 72 h at 30°C, and sampled at indicated time for measurement of cell growths and residual heavy metal concentrations by AAS (n = 3)



**Fig. 4** Measurement of residual heavy metal in biosorption assay. Biosorption of mixed heavy metals were using a sequential removal strategy. The 100 ml of soil extract was inoculated with 1  $OD_{600}$  of GYF-1 concentrate, and shaking at 30°C for 24 h. The culture was centrifuged, 5 ml of supernatant was sampled, and the rest was inoculated again with GYF-1 concentrate for another 24 h incubation at 30°C. The residual concentration of (**A**) Cd<sup>2+</sup>, (**B**) Cr<sup>6+</sup>, (**C**) Cu<sup>2+</sup>, (**D**) Ni<sup>2+</sup>, (**E**) Pb<sup>2+</sup>, (**F**) Zn<sup>2+</sup>, or Fe<sup>2+</sup> left in all collected samples were determined by AAS (n = 3)

Table 3 Predicted genes in LPS biosynthesis

Gene Tag	Gene	Annotation
KDO biosynthesis		
GYF_00028	kdsA	3-deoxy-D-manno-octulosonic acid (KDO) 8-phosphate synthase
GYF_02650	kdsB	3-deoxy-manno-octulosonate cytidylyltransferase
GYF_02059	kdsC	3-deoxy-D-manno-octulosonate 8-phosphate phosphatase
GYF_02058	kdsD	arabinose-5-phosphate isomerase
GYF_03747	atoA	3-deoxy-D-manno-octulosonate 8-phosphate phosphatase
GYF_03746	atoD	Arabinose-5-phosphate isomerase
Lipid A biosynthesis		
GYF_01039	IpxA	UDP-N-acetylglucosamine acyltransferase
GYF_01040	lpxB	Lipid A disaccharide synthetase
GYF_01706	IpxC	UDP-3-O-[3-hydroxymyristoyl] N-acetylglucosamine deacetylase
GYF_01037	lpxD	UDP-3-O-[3-hydroxymyristoyl] glucosamine N-acyltransferase
GYF_01102	lpxH	UDP-2,3-diacylglucosamine hydrolase
GYF_02844	ІрхК	Tetraacyldisaccharide-1-P 4'-kinase
GYF_03243	lpxL-1	Lipid A biosynthesis lauroyl acyltransferase
GYF_03244	lpxL-2	Lipid A biosynthesis lauroyl acyltransferase
Core oligosaccharide biosynthesis		
GYF_03850	rfaG	Glycosyltransferase
GYF_02341	rfaL	O-Antigen ligase
GYF_02561	rfaS	Lipopolysaccharide core biosynthesis protein
GYF_00179	kdtA	3-deoxy-D-manno-octulosonic-acid transferase
GYF_02340	wcaA	UDP-glucose LOS-beta-1,4 glucosyltransferase
O-polysaccharide receptor biosynthesis		
GYF_03348	wbbL	Rhamnosyltransferase
GYF_00472	rmIA1	Glucose-1-phosphate thymidylyltransferase 1
GYF_00473	rmIC	dTDP-4-dehydrorhamnose 3,5-epimerase
GYF_02287	rmID	dTDP-4-dehydrorhamnose reductase
GYF_02289	rmIB	dTDP-glucose 4,6-dehydratase
Lipid A modification		
GYF_03803	eptA	Lipid A phosphoethanolamine transferase
LPS transport		
GYF_00897	lptA	Lipopolysaccharide transport periplasmic protein
GYF_00165	lptD	Lipopolysaccharide transport protein D
GYF_00912	lptG	LPS export ABC transporter permease

On the other hand, the up-regulation of the KDO and lipid A biosynthesis was consistent with the higher relatively expressions of components in Type II and Type III secretion pathway, which are probably required for exportation of LPS synthesis-related enzymes (Fig. S8 and Table S3). Taken together, these suggest *L. aestuarii* GYF-1 increased level of lipopolysaccharides and to adaptively respond to mixed heavy metal stress.

# Biofilm formation involves in mixed heavy metals adaptation

LPS was involved in biofilm formation [19]. To verify our hypothesis if biofilm formation involves in mixed heavy metal adaptation, we first checked the cell morphology in SHMM agar plate. *L. aestuarii* GYF-1 formed clear, thick, and tight biofilm under mixed heavy metal stress compared to the vague, smooth, and loosen shape under normal LB growth (Fig. 7a and b). Second, the relative expression levels with regards to biofilm formation were checked in liquid  $1 \times$  SHMM culture (Fig. 7c). As expected, three annotated positive transcriptional regulators, GYF\_01326, GYF\_02184, GYF\_02478, and peptidylprolyl isomerase encoding gene GYF\_01776 were all up-regulated. The linear homopolymer poly-β-1,6-Nacetyl-d-glucosamine (PGA) are the gene product of the pgaABCD operon and functions as an adhesin for the maintenance of biofilm structural stability [20]. Results indicated expressions of pgaA (PGA export porin) and pgaB (PGA N-deacetylase) were at least 2-fold higher in mixed heavy metal stress that in normal growth; the expressions of two PGA synthase, pgaC and pgaD, were slightly increased but to a lesser extent. In addition, the up-regulated secretion pathway might also contribute to biofilm formation (Fig. S8). Taken together, these results indicated biofilm formation as an adaptive method

## Ctrl HM

Ctrl HM



Fig. 5 Evaluation of lipopolysaccharides in mixed heavy metal stress. LPS amounts and biofilm formation were evaluated under control and mixed heavy metals conditions. Silver staining (left) of extracted LPS, and coomassie blue R-250 staining (right) used to examine protein contamination. 10 mg of cell pellets were subjected to LPS extraction with two biological replicates, 10 µL of each sample was loaded to each lane for relative LPS level comparison

used by L. aestuarii GYF-1 for mixed heavy metal detoxification.

#### Discussions

The present study isolated one heavy metal-resistant L. aestuarii strain from soil samples near steel factory and demonstrated its potentiality in biosorption of mixed heavy metals. Currently, the Lampropedia genus is represented by four species, L. hyalina, L. puyangensis, L. cohaerens, and L. aestuarii, however, their heavy metal resistance and potential in application remain elusive. Here, we experimentally demonstrated L. aestuarii GYF-1 is resistance to multiple heavy metal stress with increased LPS level and biofilm formation.

L. aestuarii GYF-1 was isolated from soil samples containing heavy metals, which suggest strain GYF-1 have developed adaptation mechanism to mixed heavy metal stress. The annotations revealed a repertoire of metal resistance genes in L. aestuarii GYF-1 genome (Table S4), suggest they might be responsible for efflux of the heavy metals. Bacterial chromium resistance usually via chromium reductase encoded by chromium reductase and efflux system, which involves chromium transporter encoded by *chrA* and a *chrA* positive regulator *chrB* [21]. The predictions of Cr-resistance genes indicated one gene GYF\_02978 encodes chromate transporter gene chrA in Scaffold 5, however, no chrB gene was identified, suggest the ChrA functions alone in strain GYF-1. Two cupin-like domain transcription activator genes chrR were found in Scaffold 1 and Scaffold 9. For copper resistance, *vfiH* (GYF\_00340) encodes multicopper polyphenol oxidoreductase laccase was found in Scaffold 1. The copZ is a known encodes cytoplasmic copper chaperone, which was predicted in Scaffold 13 (GYF\_00913). A MerR-family transcription factor cueR was predicted in Scaffold 8 (GYF\_03485). In addition, GYF\_01534 was predicted as a potential copper(I)-binding protein although its function was not annotated in Lampropedia spp. Analysis of nickel transporter predicted many genes encode nik-ABCDE transporter in GYF-1 genome. This could explain previous finding that strain GYF-1 have high nickel tolerance but with no adsorption in nickel stress, because the main mechanism of nickel resistance used by GYF-1 is via efflux mediated by transporters. A whole nikABCDE gene cluster (GYF 01655-01659) was annotated in Scaffold 2, and partial nik gene clusters were predicted in Scaffold 3, 8, 11, 16, 20, suggest the nik transporter complex in GYF-1 might be different with those classic systems reported in model bacteria. Besides those single heavy metal specific transporter, many multiple heavy metal transporters were also annotated. The czcABC gene cluster (GYF\_02595-02597) was annotated in Scaffold 4, which encode heavy metal efflux pump of cobalt, zinc, and cadmium. Two P1B-type ATPase genes zntA (GYF\_03486 and GYF\_03587) which confer resistance specifically to Pb<sup>2+</sup>, Zn<sup>2+</sup>, and Cd<sup>2+</sup> were found in Scaffold 8. The ABC-type  $Mn^{2+}/Zn^{2+}$  transport system *znu* was predicted in Scaffold 5, with two znuB (GYF\_02916 and GYF\_02917) and one znuC (GYF\_02918), suggest the difference of znuABC complex in L. aestuarii GYF-1 from other studied species. The presence of multiple efflux transporters could also explain L aestuarii GYF-1 exhibited high tolerance to individual heavy metal stress as indicated by MIC values (Table 2). To note, the mechanism that bacterial responses to individual heavy metal or mixed stress might be different. Therefore, the role of these transporters in heavy metal stress needs to be further validated.

Polyphosphates (polyP), as polyanions, are involved in detoxification of heavy metals [22]. Annotation revealed two polyphosphate kinase encoding genes, ppk1 (GYF\_02745) and ppk2 (GYF\_02770), are involved in polyphosphate metabolic process. The expression of *ppk1* and *ppk2* was assessed to evaluate if inorganic polyP contributes to intracellular heavy metal sequestration. The expression of *ppk1* was 2.18-fold higher in



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Fig. 6 Relative expressions of LPS synthesis genes response to mixed heavy metals. Relative gene expressions involved in (a) KDO biosynthesis; (b) lipid A biosynthesis; (c) core oligosaccharide biosynthesis; (d) O-polysaccharide receptor biosynthesis. Bars represented the mean of three technical replicates and the error bars the standard error of the mean (mean ± SEM, n = 3)



Fig. 7 Evaluation of biofilm formation in mixed heavy metal stress. (a and b) cell morphologies on (a) LB agar and (b) LB agar supplemented with SHMM (n = 5). (c) Relative gene expressions involved in biofilm formation. Bars represented the mean of three technical replicates and the error bars the standard error of the mean (mean  $\pm$  SEM, n = 3)

 $1 \times$  SHMM than in control medium, and no significant up-regulation was found with regards to *ppk2* expression (Fig. S9). This indicated the polyP might contribute to mitigation of mixed heavy metal stress, but might not play the major role for biosorption. One explanation is that the intracellular compartment depends on the concentration of heavy metals inside the cells. In addition, genome annotation also predicted the presence of lipoic acid biosynthesis pathway [23] and superoxide dismutase (SOD) [24], which are important to alleviate reactive oxygen stress generated by heavy metals.

Biosorption is defined as adsorption of substances by using passive physiochemical pathways, such as electrostatic forces and ion/proton displacement, while bioaccumulation is active metabolic event in which heavy metals are taken up into the cell [5]. In this study, our current finding suggested detoxification of mixed heavy metal stress mainly through biosorption. For bioaccumulation, genome annotations indicated the presence of two proteins containing heavy metal binding motif (Table S1). GYF\_02298 was annotated as a thiol-disulfide isomerase containing CXXC motif which is responsible for multi heavy metal binding [25]. GYF 03532 was predicted as a heavy metal sensor kinase containing a sensor histidine kinase domain [26]. However, if these two proteins actively contribute to bioaccumulation of heavy metal remain elusive.

For biosorption application, we applied two-round adsorption strategy into removal of mixed heavy metals pollution. L. aestuarii GYF-1 showed biosorption efficiency of greater than 90% in removing Cd<sup>2+</sup>, Cr<sup>6+</sup>, and Pb<sup>2+</sup>, however, relatively low efficiency in adsorbing Cu<sup>2+</sup> and Zn<sup>2+</sup> and no affinity to Ni<sup>2+</sup>, individually. The biosorption efficiency is impacted by various factors, such as pH, temperature, initial concentration of the heavy metals, cell density, treatment time, etc [5]. Thus, optimization of adsorption condition might be necessary to improve the biosorption efficiency. On the other hand, mixed heavy metal biosorption assay proved the concept that L. aestuarii GYF-1 could be used as biosorbent for bioremediation of soil samples in a lab-scale. Several bioprocess factors also should be taken into considerations when scale-up to industrial level, for example type of bioreactor, pH, and temperature control, mixing and agitation, feeding strategy (in batches or in continuous mode) [27, 28].

#### Conclusions

In this study, we isolated a mixed heavy metal resistant *L. aestuarii* strain from soil samples near steel factory. Heavy metal tolerance assay indicated *L. aestuarii* GYF-1 possessed MIC value of 2 mM, 10 mM, 6 mM, 4 mM, 6 mM, 0.8 mM, and 4 mM for  $CdCl_2$ ,  $K_2CrO_4$ ,  $CuCl_2$ , NiCl<sub>2</sub>, Pb(CH<sub>3</sub>COO)<sub>2</sub>, ZnSO<sub>4</sub>, and FeCl<sub>2</sub>, respectively.

The biosorption assay demonstrated its capacity in bioremediation of soil polluted by mixed heavy metals. Genome analysis revealed abundance of heavy metal resistance genes in the genome of *L. aestuarii* GYF-1. Further evaluation on differential gene expressions under stress condition suggest enhanced LPS and biofilm formation contributed to the adaptation of mixed heavy metals. This study demonstrated *L. aestuarii* GYF-1 can be selected as a potential candidate strain for biosorption application.

#### **Materials and methods**

#### Soil sample collection and analysis

For soil sample collection, 10–20 g soils a depth of 10 cm below the surface, and a total of 10 samples were collected using an ordinary shovel. The sampling site was 3-km away from a steel factory in Liuzhou, Guangxi Province, South China. Each sampling site was spaced at least 500 m apart.

To determine the heavy metal concentration, the collected sludge samples were treated according to GB 15,618–2018 [29]. Briefly, the soil samples were air-dried and sieved. Then, 0.5 g of dry soil were digested with an acid mixture (hydrofluoric acid-nitric acid-hydrochloric acid). The digested solution was cooled, filtered, and finally diluted to 25 mL. The concentrations of heavy metal were determined by atomic absorption spectroscopy (AAS). The original concentration of Cd, Cr, Cu, Ni, Pb, Zn, and Fe ions were  $1.1\pm1.0$ ,  $65\pm17$ ,  $64\pm23$ ,  $44\pm18$ ,  $167\pm42$ ,  $262\pm26$ , and  $280\pm65$  mg/kg soil, respectively. Hence, the concentration of Cd, Cr, Cu, Ni, Pb, Zn, and Fe ions in digested solution were  $0.2\pm0.18$ ,  $25\pm6.54$ ,  $20.14\pm7.24$ ,  $15.52\pm6.35$ ,  $16.12\pm4.05$ ,  $80.15\pm7.95$ , and  $100.27\pm23.28$  µM, respectively.

#### Media and strain isolation

A stimulated heavy metal medium (SHMM) containing seven heavy metals was formulated that mimics the heavy metals that present in the soil digested solution. Thus, 1  $\times$  SHMM contains basal LB medium and supplemented with 0.2  $\mu$ M CdCl<sub>2</sub>, 25  $\mu$ M K<sub>2</sub>CrO<sub>4</sub>, 20  $\mu$ M CuSO<sub>4</sub>, 15  $\mu$ M NiCl<sub>2</sub>, 15  $\mu$ M Pb(CH<sub>3</sub>COO)<sub>2</sub>, 80  $\mu$ M ZnSO<sub>4</sub>, and 100  $\mu$ M FeSO<sub>4</sub>.

For isolation of heavy metal resistant bacteria, 0.5 g of sieved soil sample was mixed with 25 ml sterile  $ddH_2O$ , vortexed at room temperature (RT) for 10 min. One microliter of suspension was used for 10-fold serial dilution, and spread 1 ml of each diluted gradients on 0.1 × SHMM agar plate. The plates were aerobically incubated at either 30°C or 37°C for 48 h, and then individual colony was purified on 0.1 × SHMM agar plate.

#### Atomic absorption spectroscopy

The biosorption of heavy metals was determined by using atomic absorption spectroscopy. The overnight cultured bacteria were inoculated in 100 mL of LB broth supplemented with individual heavy metal or with  $1 \times SHMM$ at 30°C for indicated time. The samples were collected, cell pellets were removed by centrifugation at 7,000 rpm for 5 min. The supernatant was digested using HNO<sub>2</sub> to oxidize to a single high-valence state or converted into inorganic compounds, and digest was diluted to an appropriate concentration. The metal concentrations were determined by using Agilent AA 240FS (Agilent). The parameters including air flow, acetylene flow, characteristic wavelength, slit width, lamp current were according to the manufacturer's instructions. The removal of heavy metal was equal to the initial concentration of heavy metal in the medium subtracted by the residual heavy metal concentration in the supernatant.

#### Identification of isolate GYF-1

The isolate was cultured in Luria-Bertani (LB) medium to observe the colony morphology, and cell shape was microscopically checked. Colony PCR was directly used for amplification of 16S rDNA by universal primers (8F: 5'-AGA GTT TGA TCC TGG CTC AG-3' and 1492R: 5'-CGG TTA CCT TGT TAC GAC TT-3'). The PCR fragments were electrophorized, sent for Sanger sequencing (Sangon, Shanghai), and blasted on NCBI website. For phylogenetic analysis, the 16 S rDNA sequences of close bacterial strains were clustered by ClustalW [30], and the phylogenetic tree was constructed through Neighbor-Joining method by MEGA 7 [31].

#### Genome analysis of L. aestuarii GYF-1

The genomic DNA of GYF-1 was prepared by using Quick-DNA Kits (ZymoResearch). The genomic DNA was quantified by Qubit dsDNA HS Assay Kit (ThermoFisher). The 500 ng of genomic DNA was used for DNA library preparation by using NEB Next Ultra DNA Library Prep Kit for Illumina. The quality of the DNA library with length of ~500 bp was electrophoresed and quantified by Thermo Qubit 4.0 (ThermoFisher). Sequencing was performed on Illumina HiSeq 4000 platform with 150 bp paired-end mode.

The acquired sequence reads were subjected to quality filtering and de novo assembly using SPAdes 3.5.0 [32]. The gaps were closed by GapFiller 1.11 [33], and sequence correction was done by PrInSeS-G 1.0.0 [34]. The *in silico* DNA-DNA hybridization (DDH) using Genome-to-Genome Distance Calculator (GGDC) [15] was performed to distinguish isolate GYF-1 at species level.

For genome annotation, the coding genes and noncoding RNAs were predicted by Prokka 1.10 [35], interspersed repeats were screened by RepeatMasker [36]. Annotation was performed with the Cluster of Orthologous Groups of proteins (COG) [37], SwissProt [38], TrEMBL [39], Protein family (PFAM) [40], Conserved Domain Database (CDD) [41], Kyoto Encyclopedia of Genes and Genomes (KEGG) database [42], and NR database [43] through a BLAST+search. Gene Ontology (GO) annotation [44] was based on results from SwissProt and TrEMBL.

#### Heavy metal tolerance of L. aestuarii GYF-1

The minimum inhibitory concentration (MIC) values of *L. aestuarii* GYF-1 were tested for heavy metals, including CdCl<sub>2</sub>, CrCl<sub>3</sub>, K<sub>2</sub>CrO<sub>4</sub>, CuSO<sub>4</sub>, NiCl<sub>2</sub>, Pb(CH<sub>3</sub>COO)<sub>2</sub>, and ZnSO<sub>4</sub>. The LB medium supplemented with 0.1 M of individual heavy metal (pH was adjusted either by NaOH or HCl to 6.8) was diluted to indicated concentration. The overnight culture broth was inoculated to final OD<sub>600</sub> of 0.05. The strain was considered tolerant to a particular heavy metal concentration if OD<sub>600</sub> higher than 0.2 at 30 °C for 48 h incubation. The MIC was defined as the lowest concentration of a heavy metal inhibiting bacterial growth for 2 days.

#### Single heavy metal removal assay

For evaluating the biosorption capacity of *L. aestuarii* GYF-1, LB medium was supplemented with 0.8 mM CdCl<sub>2</sub>, 2 mM CrCl<sub>3</sub>, 2mM K<sub>2</sub>CrO<sub>4</sub>, 2mM CuSO<sub>4</sub>, 2mM NiCl<sub>2</sub>, 6 mM Pb(CH<sub>3</sub>COO)<sub>2</sub>, or 0.6 mM ZnSO<sub>4</sub>, respectively. The pH of the LB-heavy metal medium was adjusted to 6.8 either by NaOH or HCl. For biosorption assay, 100 ml of medium was inoculated with overnight culture to OD<sub>600</sub> of 0.05, and incubated at 30°C with shaking. Every 12 h, 5 ml culture was sampled for measurement of OD<sub>600</sub> and residual heavy metal concentration by AAS. The removal rate of individual heavy metal in the sample was calculated by Eq. 1.

Heavy metal removal (%) =  $(C_{init} - C_{sample}) / C_{init} \times 100$ 

where  $C_{init}$  is the initial concentration of heavy metals in the medium;  $C_{sample}$  is the concentration of heavy metal in collected sample.

#### Mixed heavy metal removal assay

The survival of *L. aestuarii* GYF-1 in mixed heavy metal stress was firstly assessed in  $1 \times$  SHMM that mimics the heavy metals that present in the soil digested solution. The overnight culture broths were inoculated at OD<sub>600</sub> of 0.05, and cultured at 30°C for 72 h. To prepare the GYF-1 concentrate used for bioremediation, the culture broth was first prepared in LB supplemented with  $1 \times$  SHMM for 24 h at 30°C, the cell suspension was washed twice

by PBS, concentrated by centrifuge, and kept at 4°C for further use.

Removal of mixed heavy metals were using a sequential removal strategy. The soil samples were air-dried and sieved. Then, 2 g of dry soil were digested with an acid mixture (hydrofluoric acid-nitric acid-hydrochloric acid). The digested solution was cooled, filtered, adjusted pH to 6.8 by NaOH and diluted to 50 mL, and then 50 mL 2  $\times$  LB medium was added for preparation of soil extract. Five milliliter of soil extracts were sampled to determine the initial heavy metal concentrations.

In first-round remediation, 1 OD<sub>600</sub> of GYF-1 concentrate was inoculated to 100 mL of soil extract, and shaking at 30°C for 24 h. The culture was centrifuged, 5 ml of supernatant was sampled, and the rest was inoculated again with GYF-1 concentrate in the same growth condition as second-round adsorption. The removal of heavy metals in collected samples were determined by AAS and calculated by Eq. 1.

#### Extraction of LPS

Extraction of LPS was using hot phenol-water method as described previously with some modifications [45]. Briefly, 10 mg of wet cell pellets were collected by centrifuge, washed twice with PBS containing 0.15 mM CaCl<sub>2</sub> and 0.5 mM MgCl<sub>2</sub>. Pellets were then resuspended in PBS, and sonicated for 10 min on ice. To eliminate the contamination of proteins and nucleic acids, 10 µg/mL of proteinase K (Roche) was added to the cell mixtures and tubes were kept at 65°C for 1 h, the mixtures were subsequently treated with 20 µg/mL of DNase I (ThermoFisher), 40 µg/mL of RNase (Sangon, Shanghai) at 37 °C overnight. Next, an equal volume of 90% hot phenol (65°C) was added to the mixtures, vortexed vigorously, and incubated at 65°C for 15 min. The mixtures were cooled down on ice, and centrifuged at 12,000 rpm for 10 min at 4°C, supernatants were transfer to new tubes, and phenol phases were re-extracted with 100 µL of water twice. The upper aqueous phase from each step were combined. The crude extracts were further precipitated by two volume of 0.375 M MgCl<sub>2</sub> in 95% ethanol at -20°C. After the samples had cooled to 0°C, they were centrifuged at 12,000 rpm for 15 min at 4°C. The pellets were re-suspended in  $1 \times$  sampling buffer, boiled at 70°C for 5 min, electrophoresed by using NuPAGE 12% Bis-Tris Gel system (ThermoFisher), and stained by either silver staining by Mass Silver Stain kit (Sangon, Shanghai) or coomassie blue R-250 (Sangon, Shanghai) according to the manufacturer's instructions.

#### **RNA** extraction

Overnight culture of strain GYF-1 was inoculated to LB with and without supplemented with  $1 \times$  SHMM at final OD<sub>600</sub> of 0.05. The cultures were incubated for 24 h

at 30°C at 200 rpm, and harvest by centrifugation at 4,000 rpm at 4°C for 10 min, and washed once by cold 1  $\times$  TE (10 mM Tris-HCl, 1 mM EDTA, pH 7.8). Extraction of total RNA was performed by using Bacteria Total RNA Isolation Kit (Sangon, Shanghai) following the manufacturer's instruction. The resuspended RNA was treated with DNase I (ThermoFisher), and purified with the RNA Clean & Concentrator kit (ZymoResearch). The resulting RNA concentration was measured with a Biospectrometer (Eppendorf).

#### cDNA synthesis and qRT-PCR

cDNA synthesis was performed using the PrimeScript RT Reagent Kit (Takara) according to the manufacturer's protocol with 1 µg of total RNA as template, and aliquot without adding PrimeScript RTase with DEPC-treated ddH<sub>2</sub>O was used as no-RT negative control to evaluate genomic DNA contamination. Primers used for qPCR were designed using Primer designing tool and list of primers was available (Table S3). The qPCR was performed using a StepOnePlus Real-Time PCR System (ThermoFisher) and the reaction mixtures were prepared in triplicate for each sample using the TB Green Premix Ex Taq (Takara). The 16 S rRNA gene was used as house-keeping gene reference. Changes in transcript abundance were automatically calculated using the  $\Delta\Delta C_T$  method.

#### Supplementary Information

The online version contains supplementary material available at https://doi.org/10.1186/s12866-023-03093-4.

Supplementary Material 1	
Supplementary Material 2	
Supplementary Material 3	
Supplementary Material 4	
Supplementary Material 5	

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Not applicable.

#### Authors' contributions

Conceptualization of the project and design: L.Y. and R.G. Strain isolation and heavy metal biosorption assay: L.Y., T.Z., J.Y., R.Z., and J.Z. Genome sequencing data analysis: F.D. and C.S. Evaluation of LPS and biofilm: L.Y. and T.Z. RNA works and qRT-PCR assessment: L.Y. T.Z., and C.S. Drafting the article or critically revising it for important intellectual content: L.Y. and R.G. Critical analysis of data and reviewing manuscript: All.

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#### **Data Availability**

Genome of *L. aestuarii* GYF-1 are available on NCBI (https://www.ncbi.nlm. nih.gov/) with accession number PRJNA958147. Results of the genome annotations and primers are available in the supplementary materials.

#### Declarations

#### Ethics approval and consent to participate

The soil samples were taken 2–3 km away from a steel factory. It's public area with no restriction for entrance and was not belonged to any industry or municipal area. Also, we are not environmental detection institution to conduct field study, the soil samples were also not polluted samples. The objective of measuring heavy metals was to set a basal heavy metal concentration for strain screening.

#### **Consent for publication**

Not applicable.

#### **Competing interests**

The authors declare no competing interests.

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